

REVIEW

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Are Northeastern U.S. forests vulnerable to extreme drought?

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Abstract

In the Northeastern U.S., drought is expected to increase in frequency over the next century, and therefore, the responses of trees to drought are important to understand. There is recent debate about whether land-use change or moisture availability is the primary driver of changes in forest species composition in this region. Some argue that fire suppression from the early twentieth century to present has resulted in an increase in shade-tolerant and pyrophobic tree species that are drought intolerant, while others suggest precipitation variability as a major driver of species composition. From this debate, an emerging hypothesis is that mesophication and increases in the abundance of mesophytic genera (e.g., *Acer*, *Betula*, and *Fagus*) resulted in forests that are more vulnerable to drought. This review examines the published literature and factors that contribute to drought vulnerability of Northeastern U.S. forests. We assessed two key concepts related to drought vulnerability, including drought tolerance (ability to survive drought) and sensitivity (short-term responses to drought), with a focus on Northeastern U.S. species. We assessed drought-tolerance classifications for species, which revealed both consistencies and inconsistencies, as well as contradictions when compared to actual observations, such as higher mortality for drought-tolerant species. Related to drought sensitivity, recent work has focused on isohydric/anisohydric regulation of leaf water potential. However, based on the review of the literature, we conclude that drought sensitivity should be viewed in terms of multiple variables, including leaf abscission, stomatal sensitivity, turgor pressure, and dynamics of non-structural carbohydrates. Genera considered drought sensitive (e.g., *Acer*, *Betula*, and *Liriodendron*) may actually be less prone to drought-induced mortality and dieback than previously considered because stomatal regulation and leaf abscission in these species are effective at preventing water potential from reaching critical thresholds during extreme drought. Independent of drought-tolerance classification, trees are prone to dieback and mortality when additional stressors are involved such as insect defoliation, calcium and magnesium deficiency, nitrogen saturation, and freeze-thaw events. Overall, our literature review shows that multiple traits associated with drought sensitivity and tolerance are important as species may rely on different mechanisms to prevent hydraulic failure and depleted carbon reserves that may lead to mortality.

Keywords: *Acer*, *Betula*, Drought sensitivity, Drought tolerance, *Quercus*, Northeastern U.S., Temperate deciduous

Review

Introduction

Northeastern U.S. forests play an important role in storing carbon (Woodbury et al. 2007) and regulating energy, biogeochemical, and hydrological climate feedbacks (Bonan 2008). In addition, temperate forest ecosystems in this region provide many economic, ecological, and recreational benefits (O'Brien 2006; Benjamin et al. 2009).

Given the high societal and ecological value of these forests, it is critical that we improve our ability to predict forest responses to future climate change. While typically considered to be light-, temperature-, and nutrient-limited (Boisvenue and Running 2006; Rennenberg and Dannenmann 2015), the historical role and potential future importance of drought in driving forest dynamics in this region is being increasingly recognized. Throughout the twentieth century, both temperature and precipitation have increased (Hayhoe et al. 2007; Wake et al. 2014), with the late twentieth century being one of the wettest periods within the last 500 years

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(Pederson et al. 2015). During this period, drought events from 1961 to 1990 typically lasted between one to 3 months and rarely exceeded 3 months (Hayhoe et al. 2007). The exception was the multi-year drought in the 1960s (1962–1965; Namias 1967, Namias 1966), which was likely the most severe drought in the last 500 years (Pederson et al. 2013). Future climate projections under high CO₂ emission scenarios indicate that more frequent droughts lasting 1–2 months throughout most of the Northeastern U.S., as well as 3–6 months in localized areas in the Northeastern U.S., will be punctuated among larger rain events (Hayhoe et al. 2007; Wake et al. 2014). The recent summer 2016 drought where large areas across the Northeastern U.S. received 25–75% of normal precipitation (July 1–September 30; <http://www.drought.gov>) may be a harbinger of such future trends.

There is growing evidence that drought influences changes in species composition at much larger time-scales (Booth et al. 2012) and reduces canopy photosynthesis and tree growth at shorter time-scales (Urbanski et al. 2007; Martin-Benito and Pederson 2015). Responses occurring at different time-scales relate to differences in species' drought tolerance, which is the ability to survive drought events, and drought sensitivity, which we define as short- to medium-term physiological response. Recent declines in mature northern hardwood forest tree species such as *Acer saccharum*, *Betula papyrifera*, and *Betula allegheniensis* (Auclair et al. 1996) and widespread dieback of *A. saccharum* (Jones and Hendershot 1989; Payette et al. 1996; Horsley et al. 2002; Roy et al. 2004; Roy et al. 2006) and *Picea rubens* (Siccama et al. 1982; Johnson 1983) have been, in part, attributed to drought along with other factors such as freeze-thaw events, insect defoliation, acid rain, nutrient availability, and forest succession. Across forests of the Northeastern U.S., tree species can vary widely in leaf photosynthetic responses to declining soil moisture availability, although species growing in mesic and wet-mesic sites tend to be more sensitive to drought than xeric sites (Kubiske and Abrams 1994). Canopy photosynthesis can be inhibited during seasonal periods of low soil moisture, reducing the overall carbon sink-strength of forests (Urbanski et al. 2007). Using dendrochronology, a number of studies have also reported reductions in tree radial growth during years of low precipitation (Cook and Jacoby 1977; Conkey 1979; Kolb and McCormick 1993; Abrams et al. 2000; Pederson et al. 2013; Martin-Benito and Pederson 2015).

The overall trend in species composition in the region has been one of “mesophication” (Nowacki and Abrams 2008), the gradual replacement of shade intolerant, fire-adapted species (*Pinus* spp., *Quercus* spp., *Carya* spp.) with shade-tolerant, fire-sensitive species (*Acer* spp., *Betula* spp., *Fagus grandifolia*). While some argue that

human influence and land use change have played the strongest role (Nowacki and Abrams 2015), others suggest that increasing precipitation over the last century is a primary driver of this tree species shift (Pederson et al. 2015). The prevailing argument for the former is that as Northeastern U.S. forests recovered from widespread forest clearing throughout the eighteenth and nineteenth centuries (Foster et al. 1998, Foster et al. 2010), fire suppression policies starting in the 1920s to present day led to a “mesophication” of central hardwoods (i.e., oak-pine forests) that involved a compositional shift towards more pyrophobic species (*Acer* spp., *F. grandifolia*, *Betula* spp., *Prunus serotina*, *Liriodendron tulipifera*, *Tsuga canadensis*) and substantial losses of pyrophilic species (*Quercus* spp., *Pinus* spp., *Castanea* spp.). However, mesophication trends during the last century have also been attributed to recent increases in precipitation based on evidence of soil moisture limitations on tree establishment, growth, and survival, as well as consistent changes among historic climate and species composition (Pederson et al. 2015). The period of 1930–2005 was one of the wettest periods on record throughout most of the eastern USA (Pederson et al. 2015), which may have contributed to the recent mesophication. Ultimately, interactions between climate and land use change have likely contributed to shifts in species composition and point to a common hypothesis: the present mesophytic species will be especially vulnerable to future severe droughts, and consequently, Northeastern U.S. Forests may experience dramatic shifts in species composition with increasing drought frequency and intensity (Abrams and Nowacki 2016). However, evidence in support of this hypothesis is largely lacking, given the historical rarity of past droughts in the region, combined with a lack of targeted research on forest response to drought—a topic not traditionally considered relevant until only recently.

Accurate understanding of species' and forest responses to drought is fundamental to making predictions of future trajectories and designing effective management approaches, as has been shown for a range of ecosystems in moisture-limited climates (Skov et al. 2004; Simonin et al. 2007; McDowell et al. 2008; Adams et al. 2012; Anderegg et al. 2015). However, similar in-depth characterization of drought-ecosystem interactions has lagged in Northeastern U.S. forests, with previous approaches largely restricted to broad classification of drought tolerance based on species' traits or growing conditions. Nevertheless, recent studies have started to examine physiological mechanisms associated with drought-sensitivity in Northeastern U.S. species (Johnson et al. 2011; Roman et al. 2015; Yi et al. 2017), thereby broadening our understanding of forest vulnerability to drought. In this review, we find that multiple ways of

characterizing drought response (e.g., tolerance and sensitivity) are important in characterizing different temporal and spatial scales. Here, we refer to drought tolerance as the ability of a particular species to survive a long-term drought event with minimal damage to branches, whereas sensitivity refers to short-term (e.g., one growing season or less), physiological responses to drought. The term vulnerability is also used in this review, which refers to longer term risk of mortality or dieback associated with drought. In this paper, we first review past approaches to classifying drought tolerance and drought sensitivity of Northeastern U.S. species, with an eye towards understanding their ability to capture consistent patterns as reflected in agreement (or inconsistency) between different studies, and between predictions and observations. Second, we review the current knowledge of linkages between drought tolerance and sensitivity to further develop our understanding of vulnerability to drought-induced mortality necessary for refining broader classifications of species drought tolerance.

Classification of drought tolerance of Northeastern U.S. species

Because drought-induced mortality events have been relatively infrequent in Northeastern U.S. forests (Allen et al. 2010), criteria including site or growing conditions, physiological responses, and plant traits have typically been used in classifying species' drought tolerance. Otherwise, drought classification would be solely based on mortality and severity of dieback. Classification of drought tolerance often relies heavily on the geographic range of species distribution (especially in relation to soil moisture availability) as an indicator of a particular species' capacity to survive moisture stress. These classification systems are developed using published sources of species' range and environmental conditions to infer drought tolerance, including the *Silvics of North America*, USDA Plants Database, USDA Tree Atlas, and US Forest Service Fire Effects Information System (Niinemets and Valladares 2006; Matthews et al. 2011; Gustafson and Sturtevant 2013; Peters et al. 2015). It is worth noting that Gustafson and Sturtevant (2013) and Niinemets and Valladares (2006) also incorporated US Forest Service Forest Inventory and Analysis (FIA) mortality data and published results of crown dieback (> 50% branch dieback) for a few species, respectively, to aid in species drought classification.

With respect to Northeastern U.S. tree species, there are both consistencies and discrepancies in how species are classified in terms of drought tolerance. Several species among studies (Niinemets and Valladares 2006; Gustafson and Sturtevant 2013; Peters et al. 2015) are consistently classified as drought-tolerant (e.g., *Quercus* spp., *P. serotina*, and *Pinus banksiana*) and drought

intolerant (e.g., *Betula* spp., *Populus* spp., *T. canadensis*, and *Abies balsamea*) (Table 1). The major inconsistencies in drought tolerance classifications occur for *Pinus strobus*, *Acer* spp., *Fraxinus americana*, *Tilia americana*, and *Picea rubens* among the three studies (Table 1). Also, drought-tolerant species (as classified by Niinemets and Valladares 2006) actually displayed a higher sensitivity (i.e., greater reductions in stem growth) during drought as compared to drought-intolerant species (Phillips et al. 2016). Likewise, species consistently classified as drought tolerant (e.g. *Quercus* spp.) experienced higher mortality following an extreme drought year in the Midwestern USA as compared with presumably more drought intolerant species, *Acer saccharum* and *Fraxinus americana* (Gu et al. 2015). The inconsistency in classification across studies presents a conundrum that plagues drought research: uncertainty in how we define tolerance and what criteria should be used to classify species in a reliable and meaningful way.

The advantage of using established classification systems is that they provide general information for nearly all tree species of the USA. A limitation of this approach is its reliance on fundamental (and often untested) assumptions about the relationship between growing conditions and vulnerability to dieback or mortality during drought (Niinemets and Valladares 2006; Gustafson and Sturtevant 2013; Peters et al. 2015). The USDA Plants Database categorizes species' drought tolerance (low, medium, high) based on field observations but not necessarily on measurements and experiments (USDA, NRCS 2017). The USDA Plants Database also does not indicate what the field observations are based on. The *Silvics of North America* provides general information about site and soil conditions (e.g., precipitation and soil texture) where species grow. However, this approach to drought tolerance classification lacks a mechanistic understanding of what determines species' tolerance, as geographic distributions may be confounded by numerous other factors (e.g., successional change, land use, competitive interactions, and biotic agents). Another potential issue with this approach is that certain species like *Tsuga canadensis* tend to grow at moist sites such as ravines, swamp edges, and north-facing slopes (Burns and Honkala 1990), which can lead to the assumption that this species is drought-intolerant. However, it is unclear whether *T. canadensis* is less tolerant of drought than other co-occurring species at the same sites. While studies have made important progress in understanding tree mortality and drought (Gustafson and Sturtevant 2013), habitat suitability of Northeastern U.S. species (Matthews et al. 2011), and linkages between shade and drought-tolerance (Niinemets and Valladares 2006), simple drought classifications may not account for specific mechanisms or functional traits that maintain growth or survival during drought.

Table 1 Drought tolerance classes for common tree species of the Northeastern U.S. as described in Niinemets and Valladares (2006), Gustafson and Sturtevant (2013), and Peters et al. (2015)

Species	Drought tolerance		
	Niinemets and Valladares (2006) (Numeric scale 1 to 5)	Gustafson and Sturtevant (2013) (No numeric scale)	Peters et al. (2015) (Numeric scale – 3 to 3)
<i>Abies balsamea</i>	Very intolerant (1.0)	Somewhat intolerant	Intolerant (– 2)
<i>Acer rubrum</i>	Intolerant (1.8)	Somewhat tolerant	Somewhat tolerant (1)
<i>Acer saccharum</i>	Intolerant (2.3)	Somewhat tolerant	Somewhat intolerant (– 1)
<i>Betula allegheniensis</i>	Somewhat tolerant (3.0)	Somewhat intolerant	Somewhat intolerant (– 1)
<i>Betula lenta</i>	Somewhat tolerant (3.0)	–	Intolerant (– 2)
<i>Betula papyrifera</i>	Intolerant (2.0)	Intolerant	Intolerant (– 2)
<i>Carya ovata</i>	Somewhat tolerant (3.0)	–	Somewhat intolerant (– 1)
^a <i>Castanea dentata</i>	Somewhat tolerant (3.0)	–	Somewhat tolerant (1)
<i>Fagus grandifolia</i>	Intolerant (1.5)	–	Somewhat intolerant (– 1)
<i>Fraxinus americana</i>	Intolerant (2.4)	Somewhat tolerant	Somewhat intolerant (– 1)
<i>Fraxinus nigra</i>	Intolerant (2.0)	Intolerant	Intolerant (– 2)
<i>Fraxinus pennsylvanica</i>	Tolerant (3.9)	–	Somewhat intolerant (– 1)
<i>Larix laricina</i>	Intolerant (2.0)	Somewhat tolerant	Somewhat tolerant (1)
<i>Picea glauca</i>	Somewhat tolerant (2.9)	Somewhat intolerant	Somewhat intolerant (– 1)
<i>Picea mariana</i>	Intolerant (2.0)	Somewhat tolerant	Intolerant (– 2)
<i>Picea rubens</i>	Somewhat tolerant (2.5)	Tolerant	Intolerant (– 2)
<i>Pinus banksiana</i>	Tolerant (4.0)	Tolerant	Tolerant (2)
<i>Pinus strobus</i>	Intolerant (2.3)	Tolerant	Very intolerant (– 3)
<i>Populus grandidentata</i>	Somewhat tolerant (2.5)	Intolerant	Intolerant (– 2)
<i>Populus tremuloides</i>	Intolerant (1.8)	Intolerant	Intolerant (– 2)
<i>Prunus serotina</i>	Somewhat tolerant (3.0)	Somewhat tolerant	Tolerant (2)
<i>Quercus alba</i>	Tolerant (3.6)	Tolerant	Somewhat tolerant (1)
<i>Quercus rubra</i>	Somewhat tolerant (2.9)	Tolerant	Somewhat tolerant (1)
<i>Quercus velutina</i>	Somewhat tolerant (3.0)	–	Tolerant (2)
<i>Thuja occidentalis</i>	Somewhat tolerant (2.7)	Somewhat intolerant	Intolerant (– 2)
<i>Tilia americana</i>	Somewhat tolerant (2.9)	Somewhat tolerant	Somewhat intolerant (– 1)
<i>Tsuga canadensis</i>	Very intolerant (1.0)	Somewhat intolerant	Intolerant (– 2)
^a <i>Ulmus americana</i>	Somewhat tolerant (2.9)	–	Intolerant (– 2)

^aSpecies that were common to the Northeastern U.S. prior to twentieth century, but are less common due to widespread mortality caused by fungal diseases. Numeric scales are listed below each citation, and scores are listed in parenthesis if provided. Tolerance classifications and scores for Niinemets and Valladares (2006) and Peters et al. (2015) were obtained from appendices, and tolerance classifications for Gustafson and Sturtevant (2013) were obtained directly from the paper (see Table 2). We revised tolerance classifications of Niinemets and Valladares (2006) by replacing the term “moderately” with “somewhat” to maintain consistency among studies.

Nowacki and Abrams (2008) proposed that the process of mesophication over the last century in Northeastern U.S. forests involves establishment and overstory dominance of shade-tolerant species that generate dense shading, moist and cool microclimates at and near the soil surface, and a relatively inflammable fuel bed. Subsequently, fire is less prone to occur in these mesophytic forest-types. If recent mesophication due to fire suppression has in fact led to greater vulnerability of Northeastern U.S. forests to drought, then this hypothesis would imply strong linkages among drought tolerance, fire tolerance, and shade intolerance across species (Abrams

and Nowacki 2016). In tree species across North America, Europe, and East Asia, there are inverse, albeit weak, correlations (R^2 of 0.02 to 0.16) between shade and drought tolerance (Niinemets and Valladares 2006). Shade-tolerant species in the Northeastern U.S. have experienced greater reductions in photosynthesis (e.g., stomatal and non-stomatal limitations) in response to drought (Kubiske et al. 1996), but declines in photosynthesis is a metric for drought-sensitivity and not necessarily associated with lower drought tolerance because declines in photosynthesis do not imply greater risk of drought-induced mortality. In other forest biomes such as

tropical dry and moist tropical forests, there is little evidence for a trade-off between drought and shade tolerance (Markesteyn and Poorter 2009). In fact, drought tolerance was more closely associated with greater biomass allocation to roots, whereas shade tolerance was more closely associated with low-specific leaf area and slower growth rates in tropical dry and moist tropical forests (Markesteyn and Poorter 2009). Less is known regarding the link between drought tolerance and pyrophilic species in temperate forests. Pyrophilic tree genera of the Northeastern U.S. generally include *Carya*, *Castanea*, *Pinus*, *Populus*, and *Quercus* (Abrams and Nowacki 2016). *Quercus* and *Castanea* spp. are considered drought-tolerant species, *Populus* spp. are mostly considered intolerant, and *Pinus* spp. are considered both tolerant and intolerant (Table 1).

Classification of drought sensitivity of Northeastern U.S. species

Plant species possess specific adaptive strategies that allow them to either avoid or permit dehydration or reductions in plant water potential in response to drought (Pallardy 2008). Species that avoid dehydration are generally considered more sensitive to drought. There are many adaptations to drought, including specific traits or responses that regulate plant water potential because water potential and components of leaf water potential, turgor pressure, and osmotic potential, are intricately linked with important physiological processes including survival, growth, gas exchange, leaf anatomy, and many biochemical processes during drought (Jones 1992; Pallardy 2008; Lambers et al. 2008).

A commonly observed adaptation to drought for Northeastern U.S. tree species is isohydric regulation of leaf water potential through stomatal control. Northern hardwood (*A. saccharum*, *B. alleghaniensis*, and *F. grandifolia*) and other species of the Northeastern U.S. display large declines in stomatal conductance with little change in leaf water potential during drought (Federer and Gee 1976; Kubiske and Abrams 1994). Species are also capable of permitting and tolerating declines in leaf water potential (e.g., anisohydry); however, this type of tolerance strategy appears to be less common. A number of studies have highlighted the anisohydric behavior of *Quercus* species compared with co-occurring species such as *Acer rubrum*, *A. saccharum*, *B. papyrifera*, *L. tulipifera*, *Populus grandidentata*, and *Sassafras albidum* (Federer 1980; Bovard et al. 2005; Yi et al. 2017). Adaptations that enable Northeastern U.S. tree species to tolerate lower water potentials generally include adjustments in osmotic potential (Parker et al. 1982; Kubiske and Abrams 1994) or cell wall thickness (Kubiske and Abrams 1994) to sustain or minimize reductions in cell turgor pressure. Within the Northeastern

U.S. there is some evidence for osmotic adjustments, particularly for species growing in more xeric sites (Kubiske and Abrams 1994). However, across a larger spatial scale, species growing in relatively wet climates, such as the Northeastern U.S., appear to be less dependent on osmotic adjustments during drought as compared with drier regions of the U.S. (Abrams 1990; Kubiske and Abrams 1992).

Since the isohydric-anisohydric framework represents an approach for classifying species drought responses and serves as an indicator of stomatal sensitivity, its definition may affect our conclusions regarding species sensitivity to drought (Table 2). Historically, species have been divided into two broad categories: isohydric or anisohydric. Isohydric species regulated water potential via stomatal closure (or reduced stomatal aperture) above a critical threshold while anisohydric species allow water potential to decline as the soil dries (Loewenstein and Pallardy 1998; Tardieu and Simonneau 1998), although the critical threshold was often ambiguous. More recently, studies have worked towards a quantitative description of the isohydric-anisohydric framework that required deriving parameters from empirical relationships of leaf water potential, stomatal conductance, and/or hydraulic conductance (Table 2, Martinez-Vilalta et al. 2014; Klein 2014; Roman et al. 2015; Skelton et al. 2015; Garcia-Forner et al. 2016; Meinzer et al. 2016). An important finding of many of these studies is that species exhibit a continuum of isohydric-anisohydric behavior rather than a dichotomy.

Despite common drought-response behavior within species, there are some inconsistencies in classification among studies (Table 2). *Quercus* species are frequently classified as displaying anisohydric behavior (Loewenstein and Pallardy 1998; Roman et al. 2015) with an exception (Martinez-Vilalta et al. 2014). This may be due, in part, to the fact that Martinez-Vilalta et al. (2014) classified species across a number of biomes (e.g., desert, temperate, and tropical), whereas Loewenstein and Pallardy (1998) and Roman et al. (2015) compared species at one site. Thus, spatial scale likely influences where species fall along the isohydric-anisohydric spectrum, with larger scales leading to a greater range between end points and, consequently, changes in species' relative position across the spectrum. Another common species of the Northeastern U.S., *A. saccharum*, was classified as isohydric (Roman et al. 2015), anisohydric (Loewenstein and Pallardy 1998), or intermediate (Klein 2014). In this case, the spatial scales were similar among two studies that reached different conclusions (Loewenstein and Pallardy 1998; Roman et al. 2015). A notable difference among these studies was that Loewenstein and Pallardy (1998) measured Ψ_{md} values at -3 MPa and greater declines in stomatal conductance, while Roman et al. (2015) measured Ψ_{md} values in the range of -0.5 to -1.0 MPa and smaller declines in stomatal

Table 2 Summary of measured and derived parameters for characterizing anisohydric and isohydric responses to drought as a proxy for drought sensitivity

Citations	Measured parameters	Derived parameters	Characteristics of isohydric/ anisohydric behavior	Related characteristics	Northeastern species
Loewenstein and Pallardy (1998)	Stomatal conductance (g_s), midday leaf water potential (ψ_{mid}), predawn leaf water potential (ψ_{pd})		Isohydric, little to no decline in ψ_{mid} with decreasing ψ_{pd} due to greater stomatal closure Anisohydric, greater decline in ψ_{mid} with decreasing ψ_{pd}	Greater concentrations of abscisic acid (ABA, root origin) during drought in isohydric species	<i>Acer saccharum</i> (A) <i>Quercus alba</i> (A) <i>Juglans nigra</i> (I)
Tardieu and Simonneau (1998)	g_s , leaf water potential (ψ_L)		Isohydric: under high evaporative demand, ψ_L reaches a plateau due to stomatal closure regardless of soil moisture conditions Anisohydric, ψ_L declines with increasing evaporative demand and declining soil moisture with midday stomatal closure under severe water deficit	Isohydric species: g_s more sensitive to ABA under high evaporative demand or water deficit. Anisohydric species: g_s shows similar response to ABA regardless of evaporative demand or soil water deficit	N/A
Sperry et al. (2002)	Midday leaf water potential (ψ_{mid}), transpiration per leaf area (E)		Isohydric, constant ψ_{mid} during high evaporative demand that is maintained above critical water potential regardless of soil moisture Anisohydric, ψ_{mid} declines with declining soil moisture until ψ_{mid} approaches a critical ψ_{mid} value	Isohydric and anisohydric species maintain smaller and larger hydraulic safety margins, respectively	N/A
Martinez-Vilalta et al. (2014)	Predawn leaf water potential (ψ_{pd}), ψ_{mid}	Slope of $\psi_{\text{pd}} - \psi_{\text{mid}}$ relationship ($\partial\psi_{\text{pd}}/\partial\psi_{\text{mid}}$)	Isohydric, $\partial\psi_{\text{pd}}/\partial\psi_{\text{mid}} = \text{zero}$ Anisohydric, $\partial\psi_{\text{pd}}/\partial\psi_{\text{mid}} > 1$	$\partial\psi_{\text{pd}}/\partial\psi_{\text{mid}}$ negatively correlated with mean summer VPD and P_{50}	<i>Quercus alba</i> (partial isohydric)
Klein (2014)	g_s , ψ_L	ψ at 50% g_s of maximum g_s ($\psi_{g_{50}}$)	Isohydric, greater $\psi_{g_{50}}$ across continuum Anisohydric, lower $\psi_{g_{50}}$ across continuum	Isohydric species have smaller P_{50}	<i>Liriodendron tulipifera</i> (I) <i>Acer saccharum</i> (intermediate)
Skelton et al. (2015)	g_s , shoot water potential (ψ), stem-specific conductivity (K_s)	ψ at 12% g_s of maximum g_s ($P_{g_{12}}$), ψ at 50% loss of conductivity (P_{50})	Isohydric, ($P_{g_{12}} - P_{50}$) = positive Anisohydric, ($P_{g_{12}} - P_{50}$) = negative	Isohydric species maintain greater hydraulic safety margin	N/A
Roman et al. (2015)	ψ_L , soil water potential (ψ_s), g_s , VPD	$\partial\psi_L/\partial\psi_s$, $\Delta\psi = \psi_L - \psi_s$	Isohydric, $\partial\psi_L/\partial\psi_s = 0$ Anisohydric, $\partial\psi_L/\partial\psi_s > 1$	Isohydric species become less sensitive to VPD due to decreasing $\Delta\psi$ and g_s during drought	<i>Quercus alba</i> (A) <i>Quercus rubra</i> (A) <i>Acer saccharum</i> (I) <i>Sassafras albidum</i> (I) <i>Liriodendron tulipifera</i> (I)
Garcia-Forner et al. (2016)	ψ_{pd} , ψ_{mid}	$\Delta\psi = \psi_{\text{pd}} - \psi_{\text{mid}}$	Isohydric, less variation in $\Delta\psi$ Anisohydric, greater variation in $\Delta\psi$	Anisohydric species showed greater stomatal control than expected	N/A
Meinzer et al. (2016)	ψ_{pd} , ψ_{mid} , leaf osmotic potential at full turgor ($\psi_{n, 100}$), Leaf water potential at turgor loss point (ψ_{TLP})	Hydroscapse area: Area of triangle bounded by $\psi_{\text{pd}} - \psi_{\text{mid}}$ regression line and 1:1 line	Isohydric, smaller hydroscapse area Anisohydric, large hydroscapse area	Hydroscapse area strongly correlated with $\psi_{n, 100}$ and ψ_{TLP} across species	N/A

Additional characteristics associated with aniso- and isohydric responses are provided. Species common to the Northeastern U.S. are also provided, as well as the type of behavior

A anisohydric, I isohydric, unless noted otherwise

conductance during the most severe portion of a drought. Yet, measurements by Roman et al. (2015) occurred during a more severe drought (June–July precipitation < 10% of mean) than the drought (May–September precipitation 53% of mean) in Loewenstein and Pallardy (1998). Consequently, both the range of the response surface and the severity of the drought can also affect evaluation of species' anisohydric-isohydric behavior.

Greenhouse experiments that grow seedlings from seeds collected from different habitats show that adaptation to local site conditions can influence how species respond to drought. Site conditions may be a critical factor contributing to the distinct differences in leaf water potential and stomatal responses to drought within species. For example, dry-site genotypes of *Cercis canadensis* (Abrams 1988), *Fraxinus americana* (Abrams et al. 1990), and *Prunus serotina* (Abrams et al. 1992) maintained greater stomatal conductance during drought despite having lower or similar leaf water potential as compared with genotypes from wetter sites. In addition to using seed sources from the Northeastern U.S., these studies also used seed sources that extended outside of the Northeastern U.S. including states like South Dakota, Nebraska (Abrams et al. 1990), Wisconsin (Abrams et al. 1992), and Kansas (Abrams 1988). Nevertheless, studies that used seed sources primarily within the Northeastern U.S. observed greater declines in stomatal conductance and photosynthesis during drought for wet-site genotypes of *Acer rubrum* (Bauerle et al. 2003) and *Quercus rubra* (Kubiske and Abrams 1992) compared to dry-site genotypes. Both studies also found that non-stomatal limitations to photosynthesis were greater in wet-site genotypes. Genotypes from drier sites also consistently had more xeromorphic leaves, smaller leaves with greater leaf mass per area and leaf thickness, despite identical environmental conditions during leaf development (Abrams 1988a; Abrams et al. 1990; Abrams et al. 1992; Kubiske and Abrams 1992). Although, both genotypic and phenotypic variation can also influence how temperate deciduous species respond to drought (Abrams 1994). For many temperate deciduous species, phenotypic variation in response to light availability strongly controls leaf and stomatal structural traits associated with drought sensitivity and can interact with genotypic variation associated with adaptation to wet or dry sites (Abrams 1994). For example, genotypic differences between dry- and wet-site seed sources were more evident in *Prunus serotina* overstory trees exposed to greater light conditions as compared with shade trees (Abrams et al. 1992). Dry-site genotypes consistently had smaller leaf area and guard cell length, as well as greater leaf thickness and leaf mass per area for sun leaves; however, there were no differences in these parameters between dry-site and wet-site genotypes for shade leaves. This has important

implications for tree response to drought and suggests that stage of development and light conditions, in addition to its topographic location across the landscape, can influence drought sensitivity.

Based on the trajectory of recent research (Table 2), characterizing species' responses to drought using the isohydric-anisohydric framework for mature trees will require intensive measurements particularly during periods of water stress. Accomplishing this for species in the Northeastern U.S. may be especially difficult, given the rarity of drought occurrence and high diversity of species in this region. A few possible solutions to this problem may include the use of experiments and targeted monitoring, as well as other measurements to determine isohydric-anisohydric behavior. Experimental work that diverts precipitation from forested plots may be one approach to identifying species sensitivity to drought while progressing trees towards dieback and mortality, which will be an important first step in understanding mechanisms associated with drought tolerance of tree species (Mencuccini et al. 2015). This approach has been implemented in a number of forested ecosystems including temperate deciduous (Tschapinski et al. 1998; Wullschlegel et al. 1998), *Eucalyptus grandis* (Battie-Laclau et al. 2014), and tropical forests (Fisher et al. 2007) in order to understand physiological responses to drought. Sap-flow measurements may circumvent logistical issues with intensive gas exchange measurements in tall, mature trees, as well as improve the timing of targeted physiological measurements. Yi et al. (2017) show that measurement of sap-flow is a reliable method for identifying species water-use strategies and stomatal sensitivity (isohydric vs. anisohydric). For example, both canopy conductance (g_c) as estimated from sap-flow measurements and stomatal conductance (g_s) as estimated from leaf gas exchange measurements displayed similar responses to drought across all species examined (Yi et al. 2017). Isohydric species (*A. saccharum*, *L. tulipifera*) experienced greater declines in both g_c and g_s than anisohydric species (*Quercus alba*, *Q. rubra*) (Roman et al. 2015; Yi et al. 2017).

Stomatal regulation of leaf water potential is just one mechanism by which plants can avoid water stress. In fact, species can avoid declines in leaf water potential through leaf abscission (Hoffmann et al. 2011) and adjustments in stomatal morphology (Abrams et al. 1994). Leaf abscission is an important strategy in temperate deciduous forests in minimizing the risk of hydraulic failure (Hoffmann et al. 2011) and aiding in recovery of photosynthesis following a drought (Gu et al. 2007) and has been reported for a number of species in the Northeastern U.S. including *A. rubrum*, *A. saccharum*, *B. papyrifera*, *Juglans nigra*, *L. tulipifera*, *Liquidambar styraciflua*, and *Nyssa sylvatica* (Federer 1980; Roberts

et al. 1980; Parker and Pallardy 1985; Pallardy 1993; Tyree et al. 1993; Gu et al. 2007; Hoffmann et al. 2011). Adjustments in stomatal morphology were found by Abrams et al. (1994) across 17 temperate deciduous species of the Northeastern U.S., where species having a smaller guard cell length maintained lower stomatal conductance and higher leaf water potential during drought. Smaller guard cell length likely increases resistance of water vapor through the stomata. Species can also postpone declines in leaf water potential during drought through stem water storage and greater rooting depth (Abrams 1990). Postponing declines in leaf water potential through stem water storage seems to be more relevant for species growing in tropical dry forests that regularly experience a dry period, as compared with temperate deciduous species (Borchert and Pockman 2005). Postponing declines in leaf water potential through greater rooting depth is more common for *Quercus* species (Abrams 1990; Pallardy 1993).

While past research on drought responses and sensitivity of Northeastern U.S. tree species primarily emphasized mechanisms associated with regulation of leaf water potential, responses that involve shifts in carbon allocation or decreases in non-structural carbohydrates (NSC; e.g., sugars and starches that support metabolic function and growth) are also indicative of drought sensitivity. In fact, models like PnET-II strongly rely on carbon allocation in modeling forest water yield and productivity in response to drought for Northeastern U.S. forests (Aber et al. 1995; Ollinger et al. 1997). For example, PnET-II estimates components of stand water balance (Aber and Federer 1992), which determines the degree of water stress and subsequent declines in net photosynthesis. PnET-II also estimates the total plant mobile carbon pool, which is dependent on net photosynthesis. The total plant mobile carbon pool is first allocated to leaf biomass, and the remaining carbon is allocated to wood production. Thus, it is assumed that carbon allocation is a passive process where mobile carbon pools simultaneously decrease with photosynthesis during drought, and consequently, leaf and wood growth is also reduced. However, there is also evidence that mobile carbon pools may be actively regulated and compete with wood growth during periods of drought or low-carbon supply (Sala et al. 2012). This hypothesis is supported in European temperate deciduous species, *Quercus petraea* and *Fagus sylvatica*, where stem growth was reduced during a summer drought, but non-structural carbohydrate (NSC) concentration did not change (Barbaroux and Bréda 2002). Similarly, stem growth of *A. saccharum* trees declined following sugar removal from xylem sap, yet sugar concentration of NSC residual pools did not change (Isselhardt et al. 2016). Thus, reduced stem growth in drought years may

be due to active regulation of mobile carbon pools at the expense of wood growth, but the relationships are complex and require further study.

Overall, drought sensitivity might be better viewed in terms of multiple variables (e.g., leaf abscission, NSC dynamics, stomatal sensitivity, rooting depth, and stomatal morphology), and not just stomatal sensitivity, that act together to prevent catastrophic losses in hydraulic functioning and carbon supply that are necessary to sustain vital physiological processes. While it is important to recognize that indicators of sensitivity vary greatly among species, adaptation to and phenotypic plasticity in response to local site conditions likely also play a critical, and perhaps even, dominant role in determining how Northeastern U.S. forests will respond to future drought conditions.

Integrating drought tolerance and sensitivity

Advancing our understanding of species' abilities to survive drought (e.g., drought tolerance) requires measurement of short-term, physiological responses to drought associated with drought sensitivity, together with their long-term response to prolonged drought to assess mechanistic relationships between drought tolerance and sensitivity. While few such integrated studies have been conducted to date, evidence thus far highlight the paradox of greater drought-induced dieback and mortality for species that are classified as drought tolerant. For example, Hoffmann et al. (2011) showed that, contrary to expectation, isohydric (e.g., drought sensitive) species with lower wood density that were less resistant to cavitation experienced less canopy dieback during a short-term drought. The higher stomatal sensitivity and leaf abscission of isohydric species (*A. rubrum*, *L. tulipifera*, *L. styraciflua*) permitted greater safety margins, while avoiding catastrophic embolism (Hoffmann et al. 2011). In contrast, anisohydric species (e.g., drought insensitive) with greater wood density (*Cornus florida*, *Q. alba*, *Q. rubra*), smaller safety margins, and an inability to shed leaves experienced greater leaf desiccation and greater canopy dieback during droughts. Thus, leaf abscission and stomatal closure are important mechanisms that reduce transpiration and minimize the risk of cavitation for temperate deciduous species (Federer 1980; Roberts et al. 1980; Lucier and Hinckley 1982; Ginter-Whitehouse et al. 1983; Gu et al. 2007). Consistent with these findings, Gu et al. (2015) observed greater mortality of drought-tolerant *Q. alba* and *Q. rubra* compared with other drought-sensitive and intolerant species (*A. saccharum*, *F. americana*) following an extreme drought in 2012, which was primarily associated with the lack of regulation of predawn water potential of the drought-tolerant species.

There are several other cases where the drought tolerant and anisohydric *Quercus* spp. experienced higher mortality following drought compared with other co-occurring temperate deciduous species in the Northeastern U.S. and other regions (Karnig and Lyford 1968; Elliott and Swank 1994; Jenkins and Pallardy 1995; Pedersen 1998; Voelker et al. 2008), which further raises questions about confidence in drought-tolerance classifications. However, there are also cases where species that tend to be classified as drought intolerant (*A. saccharum*, *F. americana*) experienced localized and regional dieback associated with drought events (Hibben and Silverborg 1978; Hendershot and Jones 1989; Roy et al. 2004). Furthermore, species classified as drought tolerant with anisohydric behavior, such as *Juniperus virginiana* (Ginter-Whitehouse et al. 1983; Bahari et al. 1985), have been shown to experience lower mortality than other co-occurring species (Gu et al. 2015).

In the Northeastern U.S., it is rare that drought acts as the sole driver of tree mortality and dieback. There are many cases in the Northeastern U.S. where mortality and dieback are associated with multiple interacting factors including drought, insect defoliators, fungal pathogens, acid rain and subsequent depletion of soil Mg and Ca, atmospheric nitrogen deposition, and soil characteristics (Karnig and Lyford 1968; Hibben and Silverborg 1978; Hendershot and Jones 1989; Pitelka and Raynal 1989; Roy et al. 2004). A number of studies have reported drought and insect defoliation or pathogen interactions that resulted in mortality and dieback in *Acer saccharum* (Kolb and McCormick 1993), *Fraxinus americana* (Hibben and Silverborg 1978), and *Quercus velutina* (Karnig and Lyford 1968).

Anthropogenic alterations in biogeochemical cycling such as acid rain and increased atmospheric nitrogen deposition in the Northeastern U.S. can also interact with drought to cause forest dieback, particularly in high elevation *Picea rubens*. Long-term acidic deposition in the Northeastern U.S. resulted in major losses of calcium and magnesium from soils (Likens et al. 1996). Calcium plays an important role in many physiological processes in trees including signal transduction during periods of stress (i.e., temperature extremes, drought, and wounding), which is critical for acclimation to environmental stress and preventing mortality (Schaberg et al. 2001). Acid rain can also directly leach calcium from *Picea rubens* leaves, which can lead to delayed responses in stomatal closure during drought (Borer et al. 2005), potentially limiting mechanisms that prevent leaf desiccation. Deposition of fine particle ammonium (NH_4^+) from the atmosphere is particularly high in high elevation sites in the Northeastern U.S. (Lovett and Lindberg 1993). Aber et al. (1989, 1998) hypothesized that nitrogen saturation may increase vulnerability of forests to drought-induced mortality because nitrogen

saturation reduces fine-root biomass while increasing leaf biomass (e.g., increase in water demand and decrease in water supply). In support of this hypothesis, McNulty et al. (2017) observed greater *Picea rubens* mortality following a few droughts in nitrogen addition plots that were part of a 30-year nitrogen saturation experiment. Furthermore, McNulty et al. (2014) hypothesized that trees growing at high nutrient sites may actually be more vulnerable to drought-induced mortality than chronically stressed trees growing at nutrient poor sites. It is likely that future drought events, coupled with nitrogen availability, may limit the resiliency of high-elevation *Picea rubens* and *Abies balsamea* forests (McNulty et al. 2017).

One reason why drought-induced mortality events are rather rare in the Northeastern U.S. may be associated with the high buffering capacity of carbon stores in temperate deciduous trees (Hoch et al. 2003; Muhr et al. 2016). Non-structural carbohydrates are not only important for sustaining growth and respiration, they also play a key role in embolism repair (Zwieniecki and Holbrook 2009; Nardini et al. 2011) and in osmotic adjustments (Wang and Stutte 1992; Guicherd et al. 1997; Hartmann and Trumbore 2016) necessary for sustaining turgor pressure under drought conditions. While characteristics of drought sensitivity such as abscission and stomatal closure may reduce overall tree carbon supply, studies have concluded that temperate deciduous trees are not carbon-limited, even during years with heavy fruiting and longer growing seasons (Hoch et al. 2003; Körner 2003). In fact, Hoch et al. (2003) showed that temperate deciduous trees contained sufficient NSCs to replace canopy leaves four times. Similarly, Muhr et al. (2016) concluded that *A. saccharum* likely contains sufficient carbon reserves to buffer against periods of low carbon supply such as defoliation or drought. Thus, carbon-limitation under short-term drought may not be an issue for drought-sensitive species that actively regulate stem and leaf water potential at the expense of reductions in carbon reserves through stomatal closure and leaf abscission.

There is considerable uncertainty regarding the ability of plants to mobilize carbon pools that are readily available for sustaining growth and metabolic processes, particularly under persistent drought. One study found that *A. rubrum* preferentially used newer NSCs for metabolism and growth, particularly in the early spring, but older NSCs (>10 yrs. old) are still accessible (Carbone et al. 2013; Richardson et al. 2013). However, not all starches may be available for mobilization (Millard et al. 2007), and drought can adversely affect long-distance transport of carbon reserves (Sala et al. 2010, 2012). In France, Bréda et al. (2006) showed that reduced starch concentration in *Q. petraea* at the end of a dry summer

resulted in greater branch dieback and incomplete budbreak and leaf-flushing the following spring; yet starch concentrations were not completely depleted. Similarly, new flushes of leaves in the spring did not occur in a declining *F. sylvatica* stand following periods of extreme waterlogging and drought despite having similar NSC concentrations in stems as compared with a healthy stand (Gérard and Bréda 2014). These results suggest that either leaf flushing is restricted at the expense of maintaining stored carbon reserves, or a proportion of stored starches within stems were not accessible during leaf-flushing. If carbon supply is in fact very large for species like *A. saccharum*, trees should be capable of using the large supply of carbon reserves following insect defoliation, but many *A. saccharum* trees in Quebec failed to produce healthy, vigorous leaves following defoliation (Roy et al. 2004). Thus, an important question related to carbon-limitation during drought is: How much of the stored NSCs are actually available for use in metabolic function during or following periods of drought?

There are many unresolved questions related to drought sensitivity and tolerance in Northeastern U.S. species. Species vulnerability to drought contributes to mortality and dieback and is ultimately determined by traits and responses associated with drought sensitivity and tolerance. However, the significance of drought vulnerability of Northeastern U.S. forests in mortality and dieback events still requires further investigation. In many cases of tree mortality and dieback, additional plant stressors such as insect defoliators, fungal pathogens, acid rain, nitrogen saturation, and soil characteristics likely interact with mechanisms associated with drought sensitivity and tolerance that exacerbate drought effects. Also, site-specific conditions (e.g., depth to bedrock, aspect, and soil drainage) that limit whole-tree carbon balance may predispose trees to mortality under drought conditions and biotic attacks. Trees growing at the fringes of their suitable habitat may be particularly vulnerable to reductions in growth and dieback (Horsley et al. 2002; Roy et al. 2004).

Future studies that examine these interactions between drought and other stressors may be fruitful as the likelihood of dieback and mortality appears to be greater when multiple stressors are involved. Past studies such as Ward et al. (2015) that have investigated multiple interacting factors including drought and nitrogen, observed greater reductions in transpiration in a combined treatment of fertilization and throughfall removal in *Pinus taeda* relative to the control and any single treatment (e.g., fertilization and throughfall removal). One suggested approach is to investigate drought responses in optimal and poor growing conditions for particular species. Many northern hardwood species like *A. saccharum*, *F. americana*, *B. allegheniensis*, and *T.*

americana have high site requirements for growth and tend to occupy well- to moderately-drained till soils dominated by calcareous bedrock, but also grow at poorer sites dominated by granite with lower nutrient availability (Leak et al. 1987). Drought experiments (e.g., throughfall removal) at sites representing both of these edaphic characteristics are one example of studying drought at optimal and poor growing conditions. A few promising responses and traits to explore include the degree of isohydry/anisohydry, leaf morphology and anatomy, propensity for osmotic adjustments, wood density, vulnerability to cavitation, and drought-deciduous behavior. More importantly, developing our understanding of critical thresholds related to soil moisture, cavitation or embolism, and low carbon reserves during extreme drought is required to improve predictive modeling for responses to drought and will help to refine habitat and species distribution modeling (Iverson et al. 2008; Iverson et al. 2017).

Conclusions

Recent mesophication and declining importance of fire-adapted and drought tolerant genera like *Quercus* may not necessarily indicate that Northeastern U.S. forests are more vulnerable to drought. Many characteristics that reflect drought sensitivity in temperate deciduous species appear to be important for surviving short-term, extreme drought. Furthermore, there are a number of cases where drought-tolerant species experienced greater dieback or mortality than other co-occurring, drought-intolerant species. This pattern does not imply though that general classification of drought-tolerance is not useful. In fact, general classification schemes that assess species' drought-tolerance using information about site conditions of where species typically grow is important, because adaptation to local site conditions strongly influences how species respond to drought. However, there is room to refine drought tolerance classification systems, as reliance on current classifications may lead to unexpected outcomes during extreme drought. Part of the solution to refining species drought-tolerance classification will likely involve future experiments that push trees past critical thresholds while measuring short-term physiological responses and specific drought-related traits for a number of species. Ideally, these experiments would include additional plant stressors (e.g., alterations in biogeochemical cycling, insect defoliation, fungal pathogens, shallow bedrock, and edaphic characteristics) to better understand potential drought interactions that lead to dieback and mortality.

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Authors' contributions

All authors contributed to the reviews and discussions on drought tolerance and sensitivity of Northeastern U.S. tree species. AC led the writing with input from all authors. All authors read and approved the final manuscript.

Competing interests

The authors declare that they have no competing interests.

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